

Methylmercury Exposure in a Subsistence Fishing Community in Lake Chapala, Mexico: An Ecological Approach

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Abstract

Background: Elevated concentrations of mercury have been documented in fish in Lake Chapala in central Mexico, an area that is home to a large subsistence fishing community. However, neither the extent of human mercury exposure nor its sources and routes have been elucidated.

Methods: Total mercury concentrations were measured in samples of fish from Lake Chapala; in sections of sediment cores from the delta of Rio Lerma, the major tributary to the lake; and in a series of suspended-particle samples collected at sites from the mouth of the Lerma to mid-Lake. A cross-sectional survey of 92 women ranging in age from 18-45 years was conducted in three communities along the Lake to investigate the relationship between fish consumption and hair mercury levels among women of child-bearing age.

Results: Highest concentrations of mercury in fish samples were found in carp (mean 0.87 ppm). Sediment data suggest a pattern of moderate ongoing contamination. Analyses of particles filtered from the water column showed highest concentrations of mercury near the mouth of the Lerma. In the human study, 27.2% of women had >1ppm hair mercury. On multivariable analysis, carp consumption and consumption of fish purchased or captured from Lake Chapala were both associated with significantly higher mean hair mercury concentrations.

Conclusions: Our preliminary data indicate that, despite a moderate level of contamination in recent sediments and suspended particulate matter, carp in Lake Chapala contain mercury levels of concern for local fish consumers. Consumption of carp appears to contribute significantly to body burden in this population. Further studies of the consequences of prenatal exposure for child neurodevelopment are being initiated.

Key Words: Methylmercury, hair mercury, fish contamination, sediment analysis

Introduction

Mercury is a ubiquitous environmental toxin. It exists in three general forms with different bioavailability and toxicity profiles -- the metallic element, inorganic mercury and organic mercury. [1] While coal-fired power plants and chloralkali plants are the leading point sources of mercury emissions in many industrialized countries, [2, 3], releases associated with amalgamation of precious metals may dominate in regions practicing artisanal mining of gold and silver. [4] Emissions from volcanoes are an important natural source and forest fires, often associated with clearing land for agriculture, can be significant [5, 6] with disproportionate impact in areas of current and recent deforestation. [7-9]

Mercury is released from combustion sources in both elemental and inorganic forms. In the atmosphere, elemental mercury is converted to inorganic ("reactive") forms that eventually deposit into soil and water. Once in natural water systems, via direct deposition or terrestrial runoff, a portion of the mercury can be transformed to an organic form, methylmercury. The microbially-mediated process occurs both in sediments and in the water column. [10]

Methylmercury is a potent neurotoxicant, especially to the developing brain. [11, 12] It biomagnifies in aquatic food chains where "nearly 100% of the mercury that bioaccumulates in upper-trophic-level fish (predator) tissue is methylmercury." [13] Highest concentrations are generally found in predatory fish at the top of the food chain - swordfish, tuna, king mackerel and shark in marine systems [14-17]; and black bass, walleye, and northern pike in freshwater systems [18]. Consumption of contaminated fish is the most important route of human exposure to methylmercury. [19] Studies in New Zealand, [20, 21] the Faroe Islands, [11, 22] and the

Seychelles Islands [23] have followed cohorts to assess the impact of fetal methylmercury exposure. In a review of these three studies, the National Academy of Sciences (NAS) found strong evidence for neurotoxicity, even at relatively low exposure. [24] Since the NAS report, an American cohort has associated elevated hair mercury levels with decreases in cognition among infants. The association persisted even when controlled for maternal fish consumption. [12]

Lake Chapala is the largest watershed in Mexico (Figure 1; Supplementary Material Table S-3) and it collects water for one-eighth of all the irrigated land in Mexico. Economic activities have created increasing pressure on the whole ecosystem prompting representatives from the concerned fishing communities to solicit an assessment of the health risk posed by industrial sewage discharges and agricultural practices (e.g., slash and burn). Previous reports of fish mercury concentrations (0.05-1.84 $\mu\text{g/g}$ wet weight), some exceeding international guidelines (0.5-1.0 ppm), have raised serious concerns about health risks to families who rely on fish from the lake for their subsistence. [25, 26]

In response to increasing community concerns about potential health risks, an environmental sampling approach was used to study pathways of methylmercury exposure in this subsistence fishing population. Mercury contamination was assessed in carp, whitefish and tilapia from Lake Chapala collected in March 2007, and contamination with polychlorinated biphenyls and persistent pesticides was explored through analysis of the fish with the highest mercury contamination. Two sediment cores and eleven suspended-particle samples were collected from the lake and analyzed for total mercury to assess the importance of historical and current inputs.

To quantify human exposure, we conducted a cross-sectional survey of 92 women aged 18-45 years old in three communities that lie on the north, west and south shores of Lake Chapala (Chapala, Jocotepec, and Tuxcueca, respectively, Figure 1) in 2008, assessing fish consumption and its correlation with mercury concentration in hair samples from women of child-bearing age.

Methods

Environmental Specimen Collection: Samples of fish from Lake Chapala, including commonly consumed carp, charales (whitefish), and tilapia were obtained by purchase from local fishermen who were observed catching the fish from the lake. The samples were originally stored at 0-4°C, but warmed to ambient temperature during the two-day transport to Rensselaer Polytechnic Institute (RPI). Subsequent testing (Supplementary Material, Table S-2) indicated that such handling does not significantly affect total mercury levels. Samples of skinless filet from tilapia and carp as well as headless, eviscerated charales, typical representations of “edible portions” of the respective species, were analyzed for total mercury using direct mercury analysis on a Milestone DMA-80 following USEPA Method 7473 (sediments, soils, and sludges). The applicability of this method to analysis of fish flesh has been previously demonstrated [27]. Additional information specific to the analyses carried out in this study is presented in the Supplementary Material (Figure S-3).

Sediment was sampled in two 28-30 cm push cores from the shallow marshy area at the east end of the lake, near the mouth of the Rio Lerma (Figure 1). Each core was divided into 2 cm sections, and maintained at room temperature throughout transport. Upon arrival at RPI, core

sections were analyzed by gamma spectroscopy to identify “recent” deposition (i.e., within the last several decades). Selected core sections were also analyzed for total mercury to determine the general level of recent sediment contamination [28].

To assess “current” (i.e. 2007) levels of total mercury on particles in Lake Chapala, eleven samples of suspended material were obtained from four different locations on the lake (Figure 1) by filtering 250 to 780 mls of water through pre-fired quartz (Whatman QMA) filters. The samples were analyzed for total mercury by direct mercury analysis at RPI. This method was developed in the laboratory of one of the authors (R.B.) and is described in the Supplementary Material.

Human Subjects and Setting: All participants were fully informed about the purposes and limitations of the study and provided a written consent in Spanish. The study protocol was approved by the Research Review Board and the Ethics Committee from the Instituto Nacional de Salud Publica (National Institute of Public Health, INSP), and the Mount Sinai School of Medicine provided L.T. exemption from review for purposes of analyzing the already collected and deidentified data.

Food frequency questionnaires that have been validated in the Mexican population [29, 30] were used to assess fish consumption. In refining the questionnaire for the Lake Chapala population, the authors identified types of fish commonly consumed in the area (tilapia, charales, carp and catfish) as well as other fish and fish soup (*sopa de pescado*), and modified the questionnaire to determine the frequency with which the participants consumed each type of fish, and the place

from which they obtained the fish (directly from the lake, purchase from a vendor who fishes from the lake, or elsewhere). Other parts of the questionnaire obtained demographic information as well as occupation, dental amalgam use, and home type.

Hair Collection: Research assistants collected a sample of approximately 50–100 strands of hair from the occipital scalp of each woman, following a well-established sample collection procedure. [31] Samples were then enclosed in individual 5 x 10 cm resealable polyethylene bags, marked with identification numbers, and maintained at room temperature while transported to RPI for analysis.

Upon receipt at RPI, each hair sample was measured and rinsed with distilled deionized water. Stainless steel scissors were used to cut 2 cm from the proximal end of each sample; these 2 cm segments were placed in glass scintillation vials which had been pre-fired at 450°C and filled with approximately 10 mL of 1% Triton X-100 solution. Following Oken et al., (2005) the vials were sonicated for 15 min., then samples were rinsed several times with distilled deionized water and dried overnight. Vials were sealed with Teflon-lined caps and held at room temperature until analysis was performed.

In preparation for mercury analysis, caps were removed and vials were covered with a Kimwipe and kept inside a plexiglas glove box for several hours while relative humidity was maintained at 55±5%. Stainless steel scissors were used to further cut 2 cm hair segments into pieces measuring 1-2 mm before transferring into quartz boats. Narrow strips cut from pre-fired quartz filter material (Whatman QMA) were pressed onto surface of samples to prevent loss during

processing. Sample weights generally ranged from 10 to 20 mg, with the goal of conducting at least two mercury analyses on the first two cm of each woman's hair.

Mercury Analysis: Analysis of total mercury content was performed by cold vapor atomic absorption spectrometry (CV-AAS) on a Direct Mercury Analyzer (DMA-80). Quality assurance was maintained by inserting a blank (an empty quartz boat) at the beginning of each sample run. Readings for blanks averaged an absorbance of .001, equivalent to ~0.04 ng of mercury. Blank readings were not significantly different from blank/blank (no sample/no boat) analyses which were run after each sample to certify that mercury carryover was negligible.

Aqueous standards and certified reference materials were included in each run, which typically contained nine samples. Certified reference materials utilized were IAEA-086 (International Atomic Energy Agency, Vienna, Austria) — cryogenically homogenized human hair with total mercury of 0.573 ± 0.039 ppm, NIES CRM-13 (National Institute for Environmental Studies, Tsukuba, Japan) — homogeneous human hair powder certified for total mercury of 4.42 ± 0.20 ppm, soil and sediment Standard Reference Materials and aqueous standards traceable to Certified Reference Solutions. The minimum quantification limit of the procedure is ~0.2 ng of total mercury, corresponding to ~20 ppb in a typical 10 mg hair sample. Additional information related to quality assurance of the hair analyses for total mercury can be found in the Supplementary Material (Table S-3).

Analysis for Persistent Organic Pollutants: The fish (a carp) with the highest mercury content was also analyzed for polychlorinated biphenyls (PCBs) and persistent pesticides as a probe for

possible contamination with these ubiquitous pollutants within the ecosystem (Supplementary Material Section B). Analyses were performed by Axys Analytical Services, Ltd. using gas chromatography/low resolution mass spectrometry method MLA-028 (PCBs as Aroclors and Chlorinated Hydrocarbon Pesticides), developed by Axys in consultation with regulatory agencies including the USEPA and NYSDEC.

Statistical Analysis: Descriptive, bivariate and multivariate analysis of hair mercury were performed, using Stata 10.0 (Stata Corporation, College Station, TX). Hair mercury concentration at the 0 -2 cm length from the base was analyzed as the dependent variable. This represents approximately two months of hair growth and a similar timescale of dietary mercury exposure (Supplementary Material, Figure S-2). Fish consumption was collected as an ordinal variable (<1 time per month, 1-3 times per month, once per week, 2-4 times per week, 5-6 times per week or daily), but, upon examining the distribution of responses, was collapsed for purposes of bivariate and multivariable analysis to <1 meal per month, at least once per month but <1 meal per week, and ≥ 1 meal per week for tilapia, whitefish, other fish and fish soup. For carp and catfish, consumption was collapsed into <1 meal per month and ≥ 1 meal per month. Source of fish was analyzed as a categorical variable (purchase from a fisherman on or near Lake Chapala or capture of fish from Lake Chapala versus purchase from a supermarket or other vendor). Amalgam number was collapsed into none, 1-2 or >2.

The distribution of hair mercury levels was found to be positively skewed, with one outlier (141 ppm, confirmed on multiple analyses). This outlier was not included in bivariate and multivariate analysis to identify predictors of hair mercury concentrations. It should be noted,

however that a similar “outlier,” a mother with 86 ppm total hair mercury, was found in a New Zealand cohort [32]. The NHANES study reported three exceptionally high hair mercury levels, all in Mexican-Americans, 849 ppm in a woman, 415.2 ppm in a one-year old child, and 109.8 ppm in a three-year old [31].

Log transformation of the remaining data resulted in a distinctly normal distribution by Kolmogorov-Smirnov test. Bivariate analysis was performed for each of the above predictor variables using generalized linear modelling with a logarithmic link to account for the skewed distribution of hair mercury. Multivariable analysis was performed using each of the fish consumption variables as purchase or capture of fish from Lake Chapala as dependent variables. Amalgam number was not included in the multivariable analysis, as it was not significantly associated with hair mercury levels on bivariate analysis.

Results

Table 1 presents mercury levels in the fish we analyzed. Among the species analyzed, carp not only have the highest levels of mercury (mean 0.87 ppm), but also the highest variability. The magnitude and variability of contamination of tilapia (mean 0.04 ppm) and whitefish (mean 0.11 ppm) were much less marked. A sample of the flesh from the carp with the highest mercury content (LLC 2, Table 2) was analyzed for PCBs and persistent pesticides. Total PCBs were 16 ppb (SDL <1 ppb), more than two orders of magnitude lower than the FDA tolerance level (2 ppm) and well within the range of “unlimited consumption, no advisory” category (< 50 ppb) of the PCB advisory for the Great Lakes. [33] The most abundant chlorinated hydrocarbons, by far,

were DDT-derived compounds. The total DDT level was 66 ppb (SDL <0.1 ppb), almost two orders of magnitude below the FDA tolerance level of 5 ppm. Furthermore, it is well below the screening value applied by the California Office of Environmental Health Hazard Assessment (100 ppb) indicating that consumption of fish contaminated at this level was judged unlikely to cause adverse health effects in humans [34].

Three samples from the upper twenty cm of sediment core Chap 2 had total mercury levels between 0.2 and 0.4 ppm dry weight (Table 3). In the 28-30 cm sample, the level was significantly lower, 0.07 ppm, a level within the range of background values reported for sediment cores [28]. Levels in core Chap 1, collected in the same general area (Figure 1), about 300 meters from Chap 2, were somewhat higher ranging from about 0.6 ppm in the upper 10 cm to a maximum of 1.3 ppm in the 18-20 cm section. The bottom section of the core, 24-26 cm, was still significantly above background at 0.6 ppm. Total mercury levels on “very recent” particles – the samples of suspended matter filtered from Lake water - ranged from 0.3 to 0.8 ppm (Table 4). The lowest levels of total mercury were seen in samples from the western most site (Alacranes, Table 4, Figure 1).

Results of hair mercury analysis are provided in Table 5. The mean hair mercury excepting the outlier was 686.4 ppb, and 27.2% of participants had hair mercury levels >1ppm. On bivariate analysis, higher carp and whitefish consumption were strongly associated with increases in hair mercury concentrations, although consumption of fish purchased from a Lake Chapala fisherman or directly captured from the Lake was not significantly associated with elevated hair mercury levels ($p=.077$). Residency in Chapala, Jocotepec or Tuxcueca was not associated with

significant differences in mean hair mercury. On multivariable analysis, consumption of fish purchased from a Lake Chapala fisherman or directly captured from the Lake was associated with 399.1 ppb higher mean hair mercury concentration (95% CI: 107.0, 856.4), while ≥ 1 meal per month carp consumption was associated with a 436.1 ppb increase (95% CI: 75.1, 1063.4). No other fish consumption variable was associated with significant differences in hair mercury levels.

Discussion

Our major findings from this pilot study of subsistence fishing population in Lake Chapala, Mexico are that (1) contamination of fish, especially carp, is significant, and (2) consumption of contaminated fish, especially carp, contributes significantly to hair mercury levels in women of childbearing age, a population of significant public health concern because of the potent prenatal neurotoxicity of methylmercury.

Levels of mercury in Lake Chapala carp were surprisingly high, even in comparison to the upper Hudson River, a system known for high levels of mercury contamination.[28] Total mercury levels in analogous samples (i.e. skinned filets) from upper Hudson carp of similar to significantly larger size ranged from 0.08 to 0.58 ppm (n = 78; median = 0.26 ppm). [18] It is likely that the high levels of mercury found in Lake Chapala carp (Table 2) are related to their particular feeding habits in the lake. It has been reported that the species of carp introduced to Lake Chapala over a hundred years ago fed extensively on whitefish eggs and fingerlings [35]. While carp are generally omnivorous, feeding significantly on plant and animal material

extracted from bottom sediment, enhanced piscivorous behavior in Lake Chapala provides a most reasonable explanation for the highly elevated levels of total mercury relative to upper Hudson carp.

The upper 10cm of sediment are generally considered the “biologically active” zone where physical and biological mixing and the feeding of benthic organisms introduce persistent contaminants to the foodchain. This is particularly true in shallow marshy areas where carp spawning has been found to effect complete mixing of the sediments to approximately this depth. [28] Our sediment cores were collected from such an area at the eastern end of Lake Chapala (Figure 1). Levels in the upper 10 cm of the cores and also in the suspended particle samples ranged from about 0.2 to 0.6 ppm, suggesting that ongoing contamination of Lake Chapala is moderate. Thirteen published studies [36] report background levels of total mercury in freshwater sediments that range from 0.01 to 0.21 ppm. For additional perspective, recent sediments from the NY/NJ Harbor deposited between the 1960s and mid 1990s had total mercury levels ranging from about 1 ppm to more than 10 ppm while less industrialized areas of the Hudson basin typically contain a few tenths of a ppm total mercury in recent deposition.[28] A study of atmospheric deposition of mercury in seven lakes located within the forest preserve of the Adirondack Mountains in NYS found levels between 0.2 and 0.6 ppm in recent sediments [36]. In our Lake Chapala cores, the level of total mercury reached approximately 1 ppm at depths between 16 and 24 cm in core Chap 1 (Table 3). This suggests that historical fluxes of mercury to the lake were significantly greater than current fluxes. Low activities of excess Pb-210 and Cs-137, the radionuclides most commonly employed in chronological studies of recent lake sediments, and the likelihood of a complex sedimentological history at the site related to

fluctuating lake levels, prevented any meaningful radiodating of these cores. The short lived natural radionuclide, Be-7, was detected in the surface (0-2 cm) section of Chap 2 indicating that the site had received at least some very recent deposition, more specifically that a significant fraction of the particles in the upper two centimeters were deposited within about a year of core collection [28]. It is planned to develop a several decades long chronology of mercury levels in Lake Chapala from sediment cores that will be collected from the deepest part of the lake, a sedimentary environment expected to have more consistent and continuous deposition with limited biological and physical mixing, conditions more amenable to radiodating.[28]

The average level of total mercury on particles filtered from Lake Chapala water collected at our western most site (345 ng/g at Alacranes, Figure 1) was more than two standard deviations lower than the mean at the other four sites (599 ± 117). This suggests further investigation of a possible source of contamination to the eastern end of Lake Chapala, perhaps associated with inputs from Rio Lerma, the largest tributary. Our technique of measuring total mercury on particles filtered from a relatively small volume of water is ideally suited to rivers such as Rio Lerma where sampling along a transect can be used to identify and evaluate point sources. In addition, there are several primary sources of mercury in the Lake Chapala basin that warrant further investigation. These include local and regional volcanism, the mining of cinnabar (HgS) in Michoacan, the Mexican state bordering the southern end of Lake Chapala, and historical artisanal gold mining in the drainage basin [37].

Our analysis of hair mercury concentrations in Lake Chapala suggests significant bioaccumulation in women of child bearing age in this population and raises the potential for

neurodevelopmental impact on their offspring. While hair mercury levels we detected were on average lower than those detected in the Faroe Islands (median 4.3 ppm) [11] and Seychelles (median 6.3 ppm) cohorts, [23] Lake Chapala women appear to have higher hair mercury levels than a US reference population (median 0.2 ppm) [31]. Our review identified only one other study of a Mexican population that assessed hair mercury. In a sample of 47 women and children from Veracruz near the Gulf of Mexico, 58% were found to have hair mercury concentrations above 1 ppm.[38] Other studies of fish and sediment mercury concentrations in Mexico have largely been performed in coastal areas near the U.S. border,[38-43] where mechanisms of contamination appear to differ from that experienced in Lake Chapala. It will also be important to assess other routes of exposure to mercury, as studies have documented the use of mercury-contaminated dietary supplements [44] as well as mercury-containing personal care products[45] and medicinal compounds.[46]

Further studies are needed to examine whether elevated hair mercury levels in the range experienced in the Lake Chapala population reflect significant risks for exposure to the developing brain, and to confirm the absence of potential confounding neurotoxicant exposures such as lead, arsenic and pesticides. The absence of “levels of concern” of PCBs in the fish with the highest mercury concentration lowers the likelihood of risks posed by this group of contaminants. Testing of human serum could rule out this confounder definitively. The usual caveats about generalizing findings from small samples are important, and these findings merit further confirmation in a larger cohort of women of childbearing age and through collection of sediment cores from the center of the lake as well as collection of additional suspended particle samples at various locations along the Rio Lerma to locate potential point sources of mercury.

While a cohort study of children born to exposed women would confirm whether this level of methylmercury exposure is problematic for neurodevelopment in this population, proactive public health guidance about the potential risk posed by consumption of carp among vulnerable populations seems warranted.

Abbreviations

Cold Vapor Atomic Absorption Spectrometry: CV-AAS

Direct Mercury Analyzer: DMA

International Atomic Energy Agency: IAEA

Instituto Nacional de Salud Publica: INSP

National Academy of Sciences: NAS

National Institute for Environmental Studies: NIES

New York State: NYS

Polychlorinated Biphenyl: PCB

Rensselaer Polytechnic Institute: RPI

LT and EC conceived the study and obtained primary funding. LT and JC collected the primary data. Analyses of hair and environmental samples were performed by MA and RFB. Statistical analysis was performed by LT, MA and RFB. LT, EC, PJJ and RFB reviewed and refined the manuscript.

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Table 1. Total mercury levels (ppm, wet weight) in fish from Lake Chapala collected in March 2007.

Species	Mean	Median	Minimum	Maximum	Number of samples
Whitefish	0.11	0.10	0.07	0.15	13
Tilapia	0.04	0.04	0.03	0.06	14
Carp	0.87	0.50	0.25	2.44	6

Table 2. Data on individual carp from Lake Chapala.

Fish ID	Source ¹	Sex	Length ² (mm)	Total Hg ³ (ppm wet)
LCC 1	Lake	Male	240	1.13±0.13
LCC 2	Lake	Female	247	2.44±0.43
LCC 3	Lake	Female	210	0.25±0.03
LCC4	Lake	Not Reported	210	0.51±0.04
LCC M1	Market	Female	180	0.39±0.04
LCC M2	Market	Not Reported	265	0.50±0.11

¹Some carp were collected directly from the lake and others were purchased at a nearby market.

²Length of the filet provided; approximately 60% of the total length of the fish.

³Values are reported with one standard deviation based on multiple analyses.

Table 3. Total mercury levels in sediment cores from the eastern end of Lake Chapala.

Core ID	Depth (cm)	Total Hg (ppm)	SD*
Chap 1	0-2	0.65	0.05
	2-4	0.69	0.05
	4-6	0.71	0.05
	6-8	0.78	0.04
	8-10	0.73	0.05
	10-12	0.82	0.06
	12-14	0.78	0.06
	14-16	0.86	0.07
	16-18	1.06	0.08
	18-20	1.28	0.10
	20-22	0.98	0.07
	22-24	1.18	0.11
	24-26	0.60	0.05
Chap 2	0-2	0.29	0.02
	8-10	0.20	0.02
	18-20	0.41	0.03
	28-30	0.07	0.01

* The relative SD based on analyses of 6-9 separate aliquots of 6 of these samples ranged from 5.7-9.7%. For the other twelve samples, the mean relative SD (7.5%) was assumed.

Table 4. Total mercury levels on suspended particles filtered from Lake Chapala in March 2007.

Site ID	Filter ID	Volume Filtered (ml)	Particle mass (g)	Total Mercury* (ng/g)
Santiago	DX 1	780	0.0272	480
Santiago	DX 2	460	0.0164	420
Lerma	DX 5	NR [#]	0.1174	670
Lerma	DX 6	250	0.0568	600
Alacranes	DY 5	530	0.0138	290
Alacranes	DY 6	225	0.0057	400
Presidio	DY 7	420	0.0101	770
Presidio	DY 8	610	0.0138	690
Petatan	DY 11	695	0.0434	580
Petatan	DY 12	340	0.0272	580

* Each filter was cut into two pieces that were analyzed separately for total mercury content. The value reported is the mean of the two measurements. In all cases except one, the relative deviation from the mean was <10%. For filter DY 7 (Presidio) it was 15%.

NR = Not Recorded.

Table 5. Hair Mercury and Fish Consumption in Women of Childbearing Age in Three Lake Chapala Communities

Population	Mean, Median Hair Mercury Concentration ppb (SD)	Percent > 1 ppm
All participants (n=92)	2218.0, 463.5 (14707.5)	27.2%
Excepting outlier (n=91)	686.4, 456.7 (720.2)	26.4%
Fish consumption, amalgam use or demographic characteristics	Mean Hair Mercury Concentration (SD), ppb	Percent > 1 ppm
Resident of Chapala (n=24)	727.7 (588.6)	29.2%
Resident of Tuxcueca (n=40)	740.0 (874.9)	20.0%
Resident of Jocotepec (n=29)	587.1 (612.9)	13.8%
Purchase fish from fisherman on Lake Chapala or catch fish from Lake Chapala (n=34)	856.0 (914.4)	23.5%
Purchase fish at supermarket, street vendor or other location (n=57)	585.3 (559.6)	12.3%
No dental amalgams (n=55)	732.2 (795.2)	20.0%
1-2 dental amalgams (n=14)	735.7 (781.3)	21.4%
>2 dental amalgams (n=22)	540.8 (435.5)	4.5%
<1 time/month carp consumption (n=47)	526.0 (536.2)	10.6%
≥1 time/month carp consumption (n=44)	857.8 (848.5)*	22.7%
<1 time/month whitefish consumption (n=49)	550.9 (474.7)	14.3%
≥1 time/month but <1 time/week whitefish consumption (n=30)	825.7 (897.8)	20.0%
≥1 time/week whitefish consumption (n=8)	1085.7 (1148.6)*	25.0%
<1 time/month fish soup consumption (n=32)	603.8 (467.8)	15.6%
≥1 time/month but <1 time/week fish soup consumption (n=33)	685.0 (619.5)	18.2%
≥1 time/week fish soup consumption (n=22)	897.7 (1100.7)	18.2%
<1 time/month tilapia consumption (n=15)	703.7 (590.0)	26.7%
≥1 time/month but <1 time/week tilapia consumption (n=33)	455.4 (452.2)	6.1%
≥1 time/week tilapia consumption (n=39)	863.0 (884.7)	20.5%
<1 time/month catfish consumption (n=52)	576.7 (656.3)	13.5%
≥1 time/month catfish consumption (n=31)	868.5 (821.0)	22.6%
<1 time/month other fish consumption (n=47)	683.5 (772.9)	14.5%
≥1 time/month other fish consumption (n=44)	695.8 (537.1)	22.7%

* p<.05, ** p<.01, *** p<.0001

Table 6. Multivariable Analysis of Hair Mercury Concentration in Women of Childbearing Age in Three Lake Chapala Communities

Fish consumption, amalgam use or demographic characteristics	Increment in Mean Hair Mercury Concentration (95% CI), ppb
Purchase fish from fisherman on Lake Chapala or catch fish from Lake Chapala	399.1 (107.0, 856.4)**
Purchase fish at supermarket, street vendor or other location	Reference
<1 time/month carp consumption	Reference
≥1 time/month carp consumption	436.1 (75.1, 1063.4)**
<1 time/month whitefish consumption	Reference
≥1 time/month but <1 time/week whitefish consumption	119.1 (-100.0, 493.5)
≥1 time/week whitefish consumption	-155.1 (-278.8, 86.3)
<1 time/month fish soup consumption	Reference
≥1 time/month but <1 time/week fish soup consumption	-74.8 (-224.9, 197.4)
≥1 time/week fish soup consumption	136.6 (-69.4, 467.6)
<1 time/month tilapia consumption	Reference
≥1 time/month but <1 time/week tilapia consumption	-224.8 (-329.0, 14.8)
≥1 time/week tilapia consumption	-155.1 (-278.8, 86.3)
<1 time/month catfish consumption	Reference
≥1 time/month catfish consumption	73.6 (-147.4, 481.2)
<1 time/month other fish consumption	Reference
≥1 time/month other fish consumption	-44.6 (-167.2, 441.6)

* p<.05, ** p<.01, *** p<.0001

Figure 1. Map of Lake Chapala and Study Locations. Suspended particle sampling sites are shown as x's and named (Table 4) for nearby rivers (Lerma and Santiago) or islands.



Figure 1

Additional files provided with this submission:

Additional file 1: Supplementary Material environmental health FINAL.doc, 68K

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